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Detection of dichloromethane with a bioluminescent (*lux*) bacterial bioreporter

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Abstract The focus of this research effort was to develop an autonomous, inducible, lux-based bioluminescent bioreporter for the real-time detection of dichloromethane. Dichloromethane (DCM), also known as methylene chloride, is a volatile organic compound and one of the most commonly used halogenated solvents in the U.S., with applications ranging from grease and paint stripping to aerosol propellants and pharmaceutical tablet coatings. Predictably, it is released into the environment where it contaminates air and water resources. Due to its classification as a probable human carcinogen, hepatic toxin, and central nervous system effector, DCM must be carefully monitored and controlled. Methods for DCM detection usually rely on analytical techniques such as solid-phase microextraction (SPME) and capillary gas chromatography or photoacoustic environmental monitors, all of which require trained personnel and/or expensive equipment. To complement conventional monitoring practices, we have created a bioreporter for the self-directed detection of DCM by taking advantage of the evolutionary adaptation of bacteria to recognize and metabolize chemical agents.

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This bioreporter, *Methylobacterium extorquens* DCM_{lux} , was engineered to contain a bioluminescent *luxCDABE* gene cassette derived from *Photorhabdus luminescens* fused downstream to the *dcm* dehalogenase operon, which causes the organism to generate visible light when exposed to DCM. We have demonstrated detection limits down to 1.0 ppm under vapor phase exposures and 0.1 ppm under liquid phase exposures with response times of 2.3 and 1.3 h, respectively, and with specificity towards DCM under relevant industrial environmental monitoring conditions.

Keywords Bioluminescence \cdot Bioreporter \cdot Dichloromethane \cdot Lux \cdot Methylene chloride

Introduction

Dichloromethane (DCM, or methylene chloride, CAS No. 75-09-2) is a colorless organic solvent widely used in paint removers and various chemical processing applications. Its industrial prevalence, low boiling point (volatility), and high water solubility make it a frequently encountered environmental contaminant. Worldwide production approaches approximately 520,000 metric tons. It is estimated that 86% of discharged DCM is released into the atmosphere where it has a half-life of 79–110 days, while the remaining 14% accumulates in water, soil, and groundwater where, it can remain for over a millennium [1, 30].

A variety of adverse health effects occur during DCM exposure. Fatal intoxication due to DCM inhalation as well as numerous cases of non-fatal poisonings by both inhalation and oral ingestion have occurred [4, 10, 15, 17, 27]. Lightheadedness, nausea, and fatigue have been reported in the workplace as acute side-effects of short-term

exposures. Extensive studies conducted in mouse, rat, and human models have provided sufficient evidence for the U.S. Environmental Protection Agency (EPA) to classify DCM as a "probable human carcinogen". More recent reevaluations have reduced the estimated risks but the mechanisms of DCM toxicity continue to remain unclear [8, 9, 28]. Once absorbed through the lungs, skin, or gastrointestinal tract, DCM is metabolized by two separate pathways, both yielding noxious by-products including carbon monoxide, carbon dioxide, hydrogen chloride, and formaldehyde [3, 9]. Among the many DCM metabolites, carbon monoxide and formaldehyde are of particular note because acutely, carbon monoxide is a known central nervous system depressant and is toxic to both the liver and cardiovascular system. Chronically, formaldehyde is a potent carcinogen. While most long-term health concerns involve the genotoxic effect of DCM metabolites, immediate and lethal DCM poisoning has been confirmed in several individuals after high-level occupational exposures [10, 15, 18]. Occupational safety standards are generally designed to minimize personal exposure through adequate ventilation. However, situations where breathing air is recycled or unventilated present cause for further precaution. For example, DCM is considered a high-priority airborne contaminant by NASA due to its known accumulation in Shuttle and International Space Station atmospheres, where Spacecraft Maximum Allowable Concentrations (SMACs) range from 100 ppm for 1 h exposures to 1 ppm over 1,000 day exposures [20]. More relevant to us here on Earth, the Occupational Safety and Health Administration (OSHA) has set permissible air concentration limits at 25 ppm as an 8-h time-weighted average (TWA) and 125 ppm as a shortterm exposure limit (STEL). The odor threshold of DCM is approximately 250 ppm. U.S. EPA safe drinking water standards are set at 5 ppm.

The detection and monitoring of workplace and environmental contaminants can be achieved via whole-cell bacterial bioreporters that sense and respond to targeted metabolites. Bioreporter bioassays for numerous contaminants have been demonstrated in water, sediment, soil, air, and wastewater environments using genetically modified bacteria typically incorporated with colorimetric, fluorescent, or bioluminescent signaling elements [24]. For the bioreporter described herein, bioluminescence derived from the bacterial lux series of genes (luxCDABE) was applied to provide real-time, reagentless signaling and assessment of target analyte bioavailability. As a first line of detection for environmental contaminants, bioluminescent bioreporters offer a rapid, inexpensive, simple, and sensitive method of monitoring, unlike traditional timeconsuming analytical techniques that require specially trained personnel and high operational cost. We developed a lux-based bioluminescent bioreporter specific to DCM by employing the evolutionary adaptation of the methylotrophic bacterium Methylobacterium extorquens (formerly M. dichloromethanicum) DM4 to scavenge for and metabolize DCM as a sole carbon source. The genes for DCM metabolism by strain DM4 are located within the dcm operon, which includes the dcmA and dcmR genes (GenBank accession #M32346) [12]. The dcmA gene encodes for a DCM dehalogenase while the dcmR gene encodes for a repressor that negatively controls the dcm operon. In the presence of DCM at low ppb concentrations, the repressor is inactivated and transcription ensues from the dcmA promoter, generating a 50-80-fold increase in dichloromethane dehalogenase [13]. Genetically engineering M. extorquens DM4 with the inducible bioluminescent reporter plasmid pCM66-dcmA/R-luxCDABE, containing a transcriptional fusion between the luxCDABE gene cassette from Photorhabdus luminescens and the dcmA/R gene of the dcm degradation operon, permitted quantifiable determination of DCM concentrations based on bioluminescence response profiles after DCM exposure.

Materials and methods

Bioreporter construction

M. extorquens strain DM4 (DSM No. 6343), a pink pigmented facultative methylotrophic bacterium, was used as the host strain for development of the DM4_{lux} bioreporter. Genetic construction involved fusion of the *dcmR/A* genetic region (GenBank accession #M32346) of strain DM4 to the P. luminescens luxCDABE gene cassette (GenBank accession #M90093) downstream to which was ligated the rrnB T_1T_2 transcriptional terminator derived from the cloning vector pKK223-3 (GenBank accession #M77749) (Fig. 1). The 1,450-bp dcmR/A region was PCR amplified from M. extorquens DM4 genomic DNA using the forward primer 5'-TCTAGACCTCCAAGGCTTGAAC-3' containing a unique XbaI site (underlined) and the reverse primer 5'-GAGCTCCACGTTATCCTCCCTT-3' containing а unique SacI site (underlined) and placed within a pCR4-TOPO vector (Invitrogen, Carlsbad, CA, USA). The luxCDABE gene cassette was PCR amplified from P. luminescens genomic DNA using the forward primer 5'-ATT AAATGGATGGCAAATAT-3' and the reverse primer 5'-GTCGACAGGATATCAACTATCAAAC-3' containing a unique SalI site (underlined) and placed within a pCR-XL-TOPO vector (Invitrogen). The rrnB transcriptional terminator was PCR amplified from pKK223-3 using the forward primer 5'-GTCGACAAGAGTTTGTAGAAAC GC-3' and the reverse primer 5'-GTCGACCTGTTT TGGCGGATG-3' each containing a SalI site (underlined) and placed within a pCR4-TOPO vector. The luxCDABE



Fig. 1 The pCM66 vector [19] was used as the backbone for construction of the dichloromethane bioreporter DCM_{*lux*}. The *dcmR*/ *dcmA* regulatory promoter region from *M. extorquens* DM4 was ligated upstream of the *P. luminescens luxCDABE* cassette (see text for specific details. *Kn* kanamycin resistance)

gene cassette was then ligated to the rrnB transcriptional terminator by cleaving the rrnB pCR4-TOPO vector with SalI to remove the rrnB sequence and ligating it into the pCR-XL-TOPO-luxCDABE vector linearized via cleavage with SalI with proper orientation confirmed by restriction mapping. This produced a pCR-XL-TOPO-luxCDABE-rrnB vector. DNA isolations were performed with Wizard Minipreps and Midipreps (Promega, Madison, WI, USA) and purified when necessary with the Geneclean Spin Kit (MP Biomedicals, Irvine, California, USA). PCR reactions were carried out in an MJ Research DNA Engine tetrad (Waltham, MA, USA) using Ready-To-Go PCR beads (Amersham Biosciences, Piscataway, NJ, USA). DNA was sequenced at all steps with the ABI Big Dye Terminator Cycle Sequencing reaction kit on an ABI 3100 DNA Genetic Analyzer (Applied Biosystems, Foster City, CA, USA).

To construct the bioreporter, each of the TOPO isolated gene sequences above were inserted into the multicloning site of the broad host range cloning vector pCM66 [19]. The *dcmR/A* genetic region was excised from its pCR4-TOPO vector with an *XbaI/SacI* double digest and directionally inserted into pCM66 similarly cleaved with *XbaI* and *SacI*. The *luxCDABE-rrnB* sequence was removed from its pCR-XL-TOPO vector via digestion with *Eco*RI and then ligated into the *dcmR/A*-pCM66 vector linearized with *Eco*RI. Proper orientation was confirmed by restriction digest mapping and sequencing. This finalized pCM66–*dcmA/R–luxCDABE* vector was electroporated into *M. extorquens* DM4 (2.5 kV, 400 Ω , 25 μ F) with transformants plated on nutrient agar plates containing 20 μ g kanamycin/ml. Exposing the plates to DCM vapor at approximately 50 ppm allowed bioluminescent colonies to be selected for using a Caliper Life Sciences IVIS imaging camera (Hopkinton, MA, USA). The brightest of these colonies, designated DCM_{lux}, was propagated for further analysis.

Growth conditions

The DCM_{*lux*} bioreporter, taken from -80° C frozen stock cultures, was maintained on nutrient agar plates containing kanamycin at 20 µg/ml to select for the lux reporter plasmid. The strain, being a slow grower, required incubation at 30°C for 4 days. Plates were refreshed from -80°C stock on a weekly basis. Liquid cultures were started from plates via inoculation into nutrient broth followed by incubation for up to 2 days at 30°C with shaking at 200 rpm to achieve an optical density at 600 nm (OD_{600}) of 0.15. Cultures were centrifuged at 10,000 rpm for 10 min and washed once in an equal volume of sterile phosphate buffered saline (PBS; in g/l, NaCl, 8; KCl, 0.2; Na₂HPO₄, 1.15; KH₂PO₄, 0.2). Cell pellets were then resuspended in 1/10th strength (0.1×) sterile minimal salts media (MSM; $1 \times \text{stock solution in g/l, KH}_2PO_4, 0.68; K_2HPO_4, 1.73;$ MgSO₄•7H₂O, 0.1; NH₄NO₃, 1.0; pH 7.0) to an OD₆₀₀ of 0.25 for application in vapor and liquid-phase bioluminescent assays.

DCM vapor phase bioluminescent assays

For vapor phase testing, DCM_{lux} cultures were immobilized in 0.7% agarose in 0.1 \times MSM. Approximate 10 ml cultures of strain DCM_{lux} were prepared in MSM at an OD_{600} of 0.25 as explained above. Liquefied 40°C agarose (2 ml) was added to 4 ml of this culture which was immediately and gently vortexed. From this mixture, 4 ml was pipetted into an 8-cm-long \times 1.5 cm² 20-ml total volume quartz flowcell. Once solidified (~ 10 min), this formed an even, immobilized layer of DCM_{lux} culture with an upper surface area that could be exposed to DCM gas as it passed through the flowcell. DCM gas was purchased from Airgas Specialty Gases, Port Allen, LA, USA at 1,000 ppm. DCM concentrations were regulated using a mass flow controller/ flow tubes (Aalborg, Orangeburg, NY, USA) and a diluting supply of Grade D breathing air (Airgas) to provide bioreporter exposures at 100, 50, 10, 1, and 0.1 ppm. The DCM/ breathing air mixture was continuously metered through sterile, chemically inert 3.1-mm-diameter Masterflex PTFE tubing (Cole Parmer Instrument Co. Chicago, IL, USA) connected to a 900-cm³ SPME glass sampling chamber,

then through the quartz flowcell positioned within the IVIS Lumina imaging chamber (Caliper Life Sciences), and finally to a waste outlet (Fig. 2). Bioluminescent imaging was monitored online in the IVIS Lumina with readings taken in photon counts per second (CPS) every 10 min over 4-h exposure durations using the integrated Living Image software package (Caliper Life Sciences) [5]. Control flowcells containing non-induced DCM_{*lux*} cultures exposed to breathing air without DCM additions were run simultaneously alongside test cultures to establish background bioluminescence. All DCM exposure experiments were performed in triplicate.

Paint stripper as a test medium

Bacterial cultures were grown and then immobilized in the quartz flowcell in the same manner as for the DCM vapor phase bioluminescent assays and exposed to dilute off-gas vapors from Crown brand Handi-Strip Semi-Paste Stripper (Packaging Service Company, Pearland, TX, USA). One milliliter of paint stripper was placed in an empty 3-l glass flask sealed with a Teflon-lined screw-cap lid. Metered breathing air was pumped continuously through an inlet located at the bottom of the flask to produce a DCM-con-taminated air supply that was directed into the 900-cm³ SPME sampling chamber, then to the flowcell in the IVIS imaging chamber, and finally, out to waste. Paint stripper exposures were performed over 4-h periods with IVIS



Fig. 2 Flow-through system for exposing the *M. extorquens* DCM_{lux} bioreporter to vapor phase DCM

images obtained every 10 min. Control flowcells containing non-induced DCM_{lux} cultures exposed to metered breathing air without paint stripper additions were run simultaneously alongside test cultures to establish background bioluminescence. All paint stripper exposure experiments were performed in triplicate.

DCM liquid-phase bioluminescent assays

A 13,000 ppm (0.15 M) stock solution of DCM was prepared by adding 490 µl of concentrated HPLC-grade DCM (Fisher Scientific, Fair Lawn, NJ, USA) to 49.5 ml sterile HPLC-grade H_2O . Cultures (100 ml) of strain DCM_{lux} were prepared in MSM at an OD₆₀₀ of 0.25 and DCM from the stock solution was added to produce initial concentrations of 100, 10, 1, 0.1, and 0.01 ppm. Cultures were immediately placed on a magnetic stirrer and continuously mixed at room temperature ($\sim 20^{\circ}$ C) while being circularly pumped at a flow rate of 340 ml/min through a sealed system consisting of the quartz flowcell positioned inside the IVIS Lumina imaging chamber. Experimental exposures were performed over 4 h with readings taken every 10 min. All associated tubing was sterile, chemically inert 3.1-mm diameter Masterflex PTFE. Control flowcells containing non-induced DCM_{lux} cultures $(0.1 \times MSM)$ without DCM) were run simultaneously alongside test cultures to calibrate for background bioluminescence. All DCM exposure experiments were performed in triplicate.

Analytical measurements of DCM concentrations

All gas-phase DCM samples were analyzed using a Hewlett Packard gas chromatograph (Model 6890) equipped with a mass spectrometer detector (MSD, Model 5973N) with an inert source. A DB-5MS column with 10-m DuraGuard $(30 \text{ m} \times 0.25 \text{ mm i.d.}, J\&W$ Scientific, Folsom, CA, USA) was used for sample separation with helium gas as the carrier gas at a constant flow rate (1.0 ml/min) maintained by an electronic pressure control module. Manual injection was used for sample analysis. The oven temperature was held at 45°C for 10 min, and then increased to a final temperature of 300°C with a 10-min hold. The injection temperature was set at 280°C and the MS source temperature was 250°C. The MS was operated in the selected ion mode (SIM) and DCM (m/z 84) was monitored. Gas-phase DCM samples were collected with a SPME fiber assembly (75 µm carboxen/polydimethylsiloxane; Supelco, Bellafonte, PA, USA). Fibers were inserted for 15 min into the 900-cm³ SPME gas-tight sampling chamber (Fig. 2) during the vapor phase bioluminescence assays (n = 3). For the liquid-phase bioluminescence assays, the SPME fibers were inserted for 15 min into the culture flask headspace (n = 3). DCM measurements were taken hourly with the SPME fibers and immediately analyzed by GC/MS [23]. SPME fibers were cleaned daily by insertion in the 280°C GC injection port for 5 min.

Liquid calibration standards were prepared at six different initial DCM concentrations ranging from 0.0013 to 130 ppm in 4.0 ml glass vials containing 2 ml of solution and sealed with a Teflon-lined septum and screw-cap lid. These standards were stirred for 30 min prior to SPME fiber exposure to assure equilibrium distribution of the DCM between the gas and liquid phases within the vials [16]. The equilibrium headspace DCM concentrations were calculated using Henry's law constant for DCM (0.00196 atm-m³/mol at 1 atm and 20°C) and plotted against the GC peak area measurements to establish a calibration curve [34]. This allowed direct measurement of the DCM exposure concentrations during the gas-phase bioluminescent assays. During the liquid-phase bioluminescence assays, the exposure concentrations were inferred using the equilibrium headspace DCM concentration and the Henry's law constant.

A factory-calibrated Innova Model 1412 photoacoustic gas monitor (LumaSense Technologies, Ballerup, Denmark) equipped with optical filter UA0980 was used to confirm the gas-phase DCM measurements made with SPME. Automatic sampling occurred every 2 min with a sample integration time of 20 s. This provided a manufacturer's specified lower detection limit of 0.05 ppm at 20°C and one atmosphere pressure. DCM concentrations measured with the Innova were maintained within 10% of the SPME vapor-phase measurements (data not shown).

Induction of *M. extorquens* DCM_{lux} by related aliphatic chlorinated industrial solvents

To characterize the specificity of the DCM_{lux} bioreporter, it was exposed to several related aliphatic chlorinated hydrocarbons commonly used as industrial solvents (trichloroethylene, tetrachloroethylene (otherwise known as perchloroethylene or PERC), 1,1,1-trichloroethane, and chloroform). Since testing each of these compounds at several different concentrations was too time-consuming using the flowcell system, they were instead tested for false-positive bioreporter induction in a higher throughput 96-well microtiter plate format in a Biotek Synergy2 microplate reader (Biotek, Winooski, VT, USA). The DCM_{lux} bioreporter was grown and resuspended in MSM to an OD₆₀₀ of 0.25 as for the liquid-phase experiments above. The first column of wells in a black 96-well microtiter plate (Dynex Technologies, Chantilly, VA, USA) contained 200 µl of bioreporter culture at a final chemical exposure concentration of 8,500 ppm. Dilutions of 1:10 were then performed throughout successive wells in 200-µl volumes to achieve a lower concentration limit of approximately 0.85 ppm. Plates were prepared on three different days using three separately grown cultures (n = 3 among experiments) with each plate containing each compound in triplicate wells (n = 3 within experiments). Each plate contained triplicate wells of strain DCM_{lux} exposed to DCM at 10 ppm (positive control) and triplicate wells of strain DCM_{lux} not exposed to a chemical to establish background bioluminescence (negative control). Plates were sealed with Breath-Easy membranes (Diversified Biotech, Boston, MA, USA) and monitored for bioluminescence in the Biotek Synergy2 at room temperature every 10 min for 24 h.

Results

Bioluminescent response kinetics of the DCM_{lux} bioreporter to vapor-phase DCM and paint-stripper volatiles

The vapor-phase bioluminescent response profile of the DCM_{lux} bioreporter was established in a flowcell format at DCM concentrations of 100, 50, 10, 1, and 0.1 ppm (actual DCM concentrations as measured by SPME (n = 3) were $101.0 \pm 0.03, 50.8 \pm 0.02, 10.5 \pm 0.05, 0.97 \pm 0.03$, and 0.14 ± 0.02 ppm). Significant bioluminescent responses (three standard deviations above control) were observed at all concentrations except 0.1 ppm and ranged from 1 h at 100 ppm to 2.3 h at 1.0 ppm (Fig. 3). Using the upper response time of 2.3 h, a linear bioluminescent response curve to increasing concentrations of DCM was generated $(R^2 = 0.99)$ (Fig. 3, inset). The paint-stripper vapor-phase DCM concentration was estimated to be 34.5 ppm using this bioluminescent response curve. SPME measurement of paint stripper volatiles at the 12.3-h time point was 39 ± 3.55 ppm. The earliest significant response during the paint stripper exposure occurred at 1.2 h.

Bioluminescent response kinetics of the DCM_{lux} bioreporter to liquid-phase DCM

A recirculating pump-driven flow-through system was used to establish the response profile for *M. extorquens* DCM_{*lux*} upon exposure to liquid-phase DCM at concentrations of 100, 10, 1, 0.1, and 0.01 ppm (DCM concentrations inferred using SPME (n = 3) were 100 ± 5.71, 50 ± 2.45, 10 ± 0.70, 1 ± 0.001, 0.1 ± 0.30, and 0.01 ± 0.07 ppm). Significant bioluminescent responses (three standard deviations above control) were observed at all concentrations except 0.01 ppm (Fig. 4). Response times ranged from 0.5 h at 100 ppm to 1.3 h at 0.1 ppm. The bioluminescent response was linear ($R^2 = 0.99$) at the upper response time of 1.3 h (Fig. 4, inset).



Fig. 3 Bioluminescent response profile in counts per second (CPS) of the *M. extorquens* DCM_{lux} bioreporter exposed to vapor-phase DCM at concentrations ranging from 1 to 100 ppm and to paint stripper. The normalized bioluminescent response was calculated by subtracting induced culture bioluminescence from background (un-exposed) culture bioluminescence for all concentrations. *Inset* Linearity of the

Induction of the DCM_{lux} bioreporter by related chlorinated solvents

M. extorquens DCM_{*lux*} was exposed in a 96-well microtiter plate to trichloroethylene, tetrachloroethylene, 1,1,1trichloroethane, and chloroform to assess the specificity of the bioreporter bioluminescent response to DCM (n = 3within and between experiments). Exposure concentrations ranged from 8,500 to 0.85 ppm. No significant bioluminescent response (three standard deviations above control) was detected in any of the samples at any of the concentrations tested over the 24-h exposure period (Fig. 5).

Discussion

The bacterial bioluminescent bioreporter *M. extorquens* DCM_{*lux*} was constructed by fusing the 1,450-bp *dcmR/ A* regulatory region from *M. extorquens* DM4 upstream of a promoterless *P. luminescens luxCDABE* gene cassette. The bioreporter proved capable of autonomously generating bioluminescence in response to vapor and liquid-phase DCM at lower tested detection limits of 1.0 and 0.1 ppm, respectively. These detection limits are well below the threshold levels for potential health risks established by

bioluminescent response to vapor-phase DCM. Points were taken from the bioluminescent profile at the upper response time of 2.3 h and plotted over concentrations ranging from 1 to 100 ppm. The *white circle* represents the 2.3-h exposure response of the DCM_{lux} bioreporter to paint stripper volatiles

OSHA for air contaminants (TWA of 25 ppm and STEL of 125 ppm) and EPA for safe drinking water (5 ppm). Response times at these lower detection limits ranged from 2.3 h under vapor-phase exposures to 1.3 h under liquid-phase exposures. Elevated DCM exposures (100 ppm) initiated bioluminescence within 1 h or less under both conditions. The parental *M. extorquens* DM4 strain can grow at an upper DCM concentration of approximately 850 ppm [14]. Thus, the bioreporter has sufficient robustness to tolerate much higher concentrations of DCM before succumbing to a toxic, disabling response. Testing at such high concentrations was not performed, however, since the purpose of the bioreporter was to signal low concentration, early warning contaminant exposures expected in work-place environments.

Volatiles emanating from enclosed commercially available paint stripper could be detected within 2.3 h at a concentration of approximately 35 ppm. This is well below the predicted DCM concentrations for consumer paint stripper applications that range from 170 to 453 ppm in residential, low-ventilation rooms [31]. Occupational exposures, for example, during aircraft paint stripping operations, produce DCM at concentrations estimated to range from 20 to 525 ppm [29, 33]. Thus, the DCM_{*lux*} bioreporter could therefore effectively pre-alert consumer and industrial workers to harmful DCM exposures.



Fig. 4 Bioluminescent response profile in counts per second (CPS) of the *M. extorquens* DCM_{lux} bioreporter exposed to liquid-phase DCM at concentrations ranging from 0.1 to 100 ppm. The normalized bioluminescent response was calculated by subtracting induced culture bioluminescence from background (un-exposed) culture



Fig. 5 Specificity of the peak bioluminescent response as fold increase above background control bioluminescence for the *M. extorquens* DCM_{*lux*} bioreporter exposed to potentially interfering liquid phase chlorinated organic solvents (trichloroethylene, tetra-chloroethylene, 1,1,1-trichloroethane, and chloroform) in a 96-well microtiter plate at final concentrations ranging from 8,500 to 0.85 ppm

No significant bioluminescence induction was observed from the DCM_{*lux*} bioreporter upon exposure to other common industrial solvents (trichloroethylene, tetrachloroethylene, 1,1,1-trichloroethane, and chloroform), thus confirming its specificity for DCM and validating minimal false-positive signaling under the tested environmental

bioluminescence for all concentrations. *Inset* Linearity of the bioluminescent response to liquid-phase DCM. Points were taken from the bioluminescent profile at the upper response time of 1.3 h and plotted over concentrations ranging from 0.1 to 100 ppm

conditions. *Hyphomicrobium* sp. DM2 (ATCC #43129), which also harbors a *dcmA* DCM dehalogenase gene [21], has been applied as a reporter in its native state in a multi-transducer flow-calorimeter/chloride-sensitive electrode biosensor format, with a detection limit of 10 ppb in water [13]. It demonstrated similar specificity but was predictively cross-sensitive to dihalomethanes besides DCM, such as bromochloromethane and dibromomethane. However, the minimal to nonexistent industrial and commercial application of these other dichloromethanes makes the DCM_{*lux*} bioreporter DCM specific within its intended monitoring environments.

In its current state, the DCM_{*lux*} bioreporter could be applied as a sensor within bioreactor treatment schemes designed for biological removal of DCM from air- and liquid-flow streams. Biotrickling filters, bioscrubbers, and similar bioreactor fabrications containing biodegradative microorganisms such as *M. extorquens* DM4 and *Hyphomicrobium* sp. DM2 have been shown to effectively reduce DCM concentrations in polluted air streams and industrial and municipal wastewater effluents [2, 11, 22]. The integration of DCM_{*lux*} bioreporters into bioreactors would serve not only to degrade DCM but to additionally report on its bioavailability, as has been previously demonstrated with, for example, the bioreporter *Pseudomonas putida* TVA8 and its targeted degradation and sensing of toluene in a packed bed reactor [26]. The measurement of resulting bioluminescent signals within the bioreactor is then typically achieved via the connection of fiber optic cables that terminate to a photomultiplier tube (PMT) or direct connection of the PMT or other light gathering device to the bioreactor itself [7]. Smaller-scale chip-based biosensors can also be functionally applied for target chemical sensing using bioreporter interfaced microelectronic circuitry. Integrated circuit microluminometers and avalanche photodiodes have both been mated with living bioluminescent bioreporters to produce self-contained biosensors on a platform of only a few square millimeters [6, 25, 32]. The potential exists to similarly assimilate DCM_{lux} bioreporters into analogous on-chip detection schemes. The ability to effectively encapsulate these bioreporters into agar-based matrices and exploiting their slow growth rate to maximize longer-term viability attests to such potential suitability.

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References

- Agency for Toxic Substances, Disease Registry (ATSDR) (2005) Toxicological profile for methylene chloride. Public Health Services, U.S. Department of Health and Human Services, Atlanta
- Bailon L, Nikolausz M, Kastner M, Veiga MC, Kennes C (2009) Removal of dichloromethane from waste gases in one- and twoliquid-phase stirred tank bioreactors and biotrickling filters. Water Res 43:11–20
- Casanova M, Bell DA, Heck HD (1997) Dichloromethane metabolism to formaldehyde and reaction of formaldehyde with nucleic acids in hepatocytes of rodents and humans with and without glutathione S-transferase T1 and M1 genes. Fundam Appl Toxicol 37:168–180
- Chang YL, Yang CC, Deng JF, Ger J, Tsai WJ, Wu ML, Liaw HC, Liaw SJ (1999) Diverse manifestations of oral methylene chloride poisoning: report of 6 cases. J Toxicol Clin Toxicol 37:497–504
- Creton R, Jaffe LF (2001) Chemiluminescence microscopy as a tool in biomedical research. Biotechniques 31:1098–1105
- Daniel R, Almog R, Ron A, Belkin S, Diamand YS (2008) Modeling and measurement of a whole-cell bioluminescent biosensor based on a single photon avalanche diode. Biosens Bioelectron 24:882–887
- Eltzov E, Marks RS (2010) Fiber-optic based cell sensors. In: Scheper T (ed) Whole cell sensing systems I: reporter cells and devices, advances in biochemical engineering/biotechnology, vol 117. Springer, Berlin Heidelberg New York, pp 131–154
- Evans MV, Caldwell JC (2010) Evaluation of two different metabolic hypotheses for dichloromethane toxicity using physiologically based pharmacokinetic modeling for in vivo inhalation gas uptake data exposure in female B6C3F1 mice. Toxicol Appl Pharmacol 244:280–290
- Fan AM, Alexeeff GV (2000) Public health goal for dichloromethane (methylene chloride, DCM) in drinking water. Office of Environmental Health Hazard Assessment and California Environmental Protection Agency. www.oehha.ca.gov

- Fechner G, Ortmann C, Du Chesne A, Kohler H (2001) Fatal intoxication due to excessive dichloromethane inhalation. Forensic Sci Int 122:69–72
- Galli R (1987) Biodegradation of dichloromethane in waste-water using a fluidized-bed bioreactor. Appl Microbiol Biotechnol 27:206–213
- Galli R, Leisinger T (1988) Plasmid analysis and cloning of the dichloromethane utilization genes of *Methylobacterium* sp. DM4. J Gen Microbiol 134:943–952
- Henrysson T, Mattiasson B (1993) A microbial biosensor system for dihalomethanes. Biodegradation 4:101–105
- Kayser MF, Stumpp MT, Vuilleumier S (2000) DNA polymerase I is essential for growth of *Methylobacterium dichloromethanicum* DM4 with dichloromethane. J Bacteriol 182:5433–5439
- Kim NY, Park SW, Suh JK (1996) Two fatal cases of dichloromethane or chloroform poisoning. J Forensic Sci 41:527–529
- Li K, Santilli A, Goldthorp M, Whiticar S, Lambert P, Fingas M (2001) Solvent vapour monitoring in work space by solid phase micro extraction. J Hazard Mater 83:83–91
- Mahmud M, Kales SN (1999) Methylene chloride poisoning in a cabinet worker. Environ Health Perspect 107:769–772
- Manno M, Rugge M, Cocheo V (1992) Double fatal inhalation of dichloromethane. Hum Exp Toxicol 11:540–545
- Marx CJ, Lidstrom ME (2001) Development of improved versatile broad-host-range vectors for use in methylotrophs and other Gram-negative bacteria. Microbiology (UK) 147:2065– 2075
- National Research Council (2008) Spacecraft maximum allowable concentrations for selected airborne contaminants. National Academy of Sciences, Washington, DC. http://www.nap.edu/ catalog/12529.html
- Nikolausz M, Kappelmeyer U, Nijenhuis I, Ziller K, Kastner M (2005) Molecular characterization of dichloromethane-degrading *Hyphomicrobium* strains using 16S rDNA and DCM dehalogenase gene sequences. Syst Appl Microbiol 28:582–587
- Osuna MB, Sipma J, Emanuelsson MAE, Carvalho MF, Castro PML (2008) Biodegradation of 2-fluorobenzoate and dichloromethane under simultaneous and sequential alternating pollutant feeding. Water Res 42:3857–3869
- Poli D, Manini P, Andreoli R, Franchini I, Mutti A (2005) Determination of dichloromethane, trichloroethylene and perchloroethylene in urine samples by headspace solid phase microextraction gas chromatography-mass spectrometry. J Chromatogr B 820: 95–102
- Ripp S, DiClaudio ML, Sayler GS (2010) Biosensors as environmental monitors. In: Mitchell R, Gu JD (eds) Environmental microbiology, 2nd edn. Wiley, New York, pp 213–233
- 25. Shacham-Diamand Y, Belkin S, Rishpon J, Elad T, Melamed S, Biran A, Yagur-Kroll S, Almog R, Daniel R, Ben-Yoav H, Rabner A, Vernick S, Elman N, Popovtzer R (2010) Optical and electrical interfacing technologies for living cell bio-chips. Curr Pharm Biotechnol 11:376–383
- 26. Shingleton JT, Applegate BA, Baker AJ, Sayler GS, Bienkowski PR (2001) Quantification of toluene dioxygenase induction and kinetic modeling of TCE cometabolism by *Pseudomonas putida* TVA8. Biotechnol Bioeng 76:341–350
- Shusterman D, Quinlan P, Lowengart R, Cone J (1990) Methylene chloride intoxication in a furniture refinisher—a comparison of exposure estimates utilizing workplace air sampling and blood carboxyhemoglobin measurements. J Occup Environ Med 32:451–454
- Starr TB, Matanoski G, Anders MW, Andersen ME (2006) Workshop overview: reassessment of the cancer risk of dichloromethane in humans. Toxicol Sci 91:20–28
- 29. Uang SN, Shih TS, Chang CH, Chang SM, Tsai CJ, Deshpande CG (2006) Exposure assessment of organic solvents for aircraft

paint stripping and spraying workers. Sci Total Environ 356:38-44

- U.S. Environmental Protection Agency (1994) Chemical summary for methylene chloride (dichloromethane). Office of Pollution Prevention and Toxics, Report number 749-F-94-018a
- van Veen MP, Fortezza F, Spaans E, Mensinga TT (2002) Nonprofessional paint stripping, model prediction and experimental validation of indoor dichloromethane levels. Indoor Air 12:92–97
- 32. Vijayaraghavn R, Islam SK, Zhang M, Ripp S, Caylor S, Weathers B, Moser S, Terry S, Blalock B, Sayler GS (2007) A

bioreporter bioluminescent integrated circuit for very low-level chemical sensing in both gas and liquid environments. Sens Actuator B Chem 123:922–928

- 33. Vincent R, Poirot P, Subra I, Rieger B, Cicolella A (1994) Occupational exposure to organic solvents during paint stripping and painting operations in the aeronautical industry. Int Arch Occup Environ Health 65:377–380
- 34. Washington JW (1996) Gas partitioning of dissolved volatile organic compounds in the vadose zone: principles, temperature effects and literature review. Ground Water 34:709–718